

for

2012

Jacobs River Estuary

Fine Scale Monitoring of Highly Eutrophic Arms 2011/2012

Cover Photo: Pourakino Arm, Jacobs River Estuary.

 Aparima Arm - Jacobs River Estuary.

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Fine Scale Monitoring of Highly Eutrophic Arms 2011/2012

Prepared for Environment Southland

By

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JACOBS RIVER ESTUARY - EXECUTIVE SUMMARY

This report summarises the results of the baseline 2012 fine scale monitoring of two eutrophic, poorly flushed, intertidal sites (Pourakino Arm and Aparima Arm) within Jacobs River Estuary, a medium-sized "tidal lagoon" type estuary (area 720ha), discharging to the sea at Riverton. It is one of the key estuaries in Environment Southland's (ES's) long-term coastal monitoring programme. The following sections summarise monitoring results, condition ratings, overall estuary condition, and monitoring and management recommendations.

Fine Scale Results

- • The sites were dominated by mud, and were poorly oxygenated (RPD was at the surface).
- • The rate of sedimentation (infilling with mud) was in the low-moderate category in 2011-2012.
- The invertebrate community was dominated by surface feeding, mud and organic enrichment tolerant species, living on the surface macroalgal layer. Very few species were present within the underlying anoxic and sulphide-rich muds.
- • Sediment nutrients and organic carbon were moderately elevated, and heavy metals were below the ANZECC (2000) ISQG-Low trigger values (i.e. low toxicity), and similar to those measured at the other fine scale sites in Jacobs River Estuary.

ESTUARY CONDITION AND ISSUES

In relation to the key issues addressed by the fine scale monitoring (i.e. sedimentation, eutrophication, and toxicity), the 2012 results indicate that the poorly flushed Pourakino Arm, and the Northern Flats of the Aparima Arm are excessively muddy, have elevated nutrients and nuisance macroalgal growths, and very poor sediment oxygenation. As a result, the macro-invertebrate community is severely degraded with little animal life able to establish in the underlying sediments, while surface feeding species are few in number and limited to those tolerant of poor conditions. Such conditions limit the food availability for fish and birdlife, and mean the ability of the estuary to assimilate nutrient and sediment loads from the catchment is exceeded. Toxicity (indicated by heavy metals) was low and similar to other sites in the estuary.

RECOMMENDED MONI TORING AN D MANA GEMENT

Eutrophication and sedimentation have been identified as major issues in Jacobs River Estuary since at least 2007-2008 (Robertson and Stevens 2008) as has been the case for several other Southland estuaries (e.g. New River Estuary, Waimatuku Estuary, and Waituna Lagoon).

To address these issues, it is recommended that catchment nutrient and sediment guideline criteria be developed for each estuary type in Southland in a prioritised fashion, with Jacobs River Estuary as the second priority behind New River Estuary. Assessing the extent to which current catchment loads meet guideline criteria will enable ES to sustainably manage the estuary and its surroundings. If the approach is followed and successfully executed, the estuary will flourish and provide sustainable human use and ecological values in the long term. If catchment loads exceed the estuary's assimilative capacity, it will continue to degrade.

In order to assess ongoing trends in the fine scale condition of the estuary it is recommended that the newly established eutrophic sites be monitored in Feb. 2013, 2014 and again in Feb. 2016. This will coincide with when the 5 yearly fine scale trend monitoring at the three existing central basin sites falls due. Broad scale sedimentation rate, seagrass, and macroalgal monitoring should continue annually, and broad scale mapping every 5 years (next due in 2013).

All photos by Wriggle except where noted otherwise.

1. In tro duction

Developing an understanding of the condition and risks to coastal and estuarine habitats is critical to the management of biological resources. The process used for estuary monitoring and management by Environment Southland (ES) in Jacobs River Estuary consists of three components developed from the National Estuary Monitoring Protocol (NEMP) (Robertson et al. 2002):

- **1. Ecological Vulnerability Assessment** (EVA) of the estuary to major issues (Table 1) and appropriate monitoring design. A preliminary EVA has been completed for Jacobs River Estuary and is reported on in Robertson and Stevens (2008).
- **2. Broad Scale Habitat Mapping** (NEMP approach). This component, which documents the key habitats within the estuary, and changes to these habitats over time, was undertaken in 2003 (Robertson et al. 2003), and repeated in 2008 (Stevens and Robertson 2008).
- **3. Fine Scale Monitoring** (NEMP approach). Monitoring of physical, chemical and biological indicators (Table 2) including sedimentation plate monitoring (established in 2011). This component, which provides detailed information on the condition of the Jacobs River Estuary, has been undertaken in 2003, 2004, 2005, 2006 (Robertson and Stevens 2006), and 2011 (Robertson and Stevens 2011).

In addition, a series of condition ratings have been developed to help evaluate overall estuary condition and decide on appropriate monitoring and management actions. These ratings, described in Section 2, currently trigger annual monitoring of sedimentation rate and macroalgal growth in the estuary.

The results of the recent annual broad scale macroalgal monitoring (Stevens and Robertson 2010, 2011, 2012), in conjunction with the monitoring undertaken from 2002-2011, has highlighted the presence of extensive and increasing eutrophication and sedimentation problems in the natural settling areas within the Aparima and Pourakino Arms of Jacobs River Estuary. The increased eutrophication symptoms (very low sediment oxygenation and sulphide-rich sediments, smothering macroalgae, rapid soft mud accumulation) correlate with increased catchment nutrient loads over the last 10 years (see ES 2012).

These symptoms have not been as conspicuous in the fine scale monitoring results to date because the sites are located on the relatively well flushed sandy intertidal flats of the estuary - the dominant habitat type in the estuary which was the focus of the original NEMP sampling design (see Figure 1, Robertson and Stevens 2011).

Therefore, in response to the eutrophication and sedimentation problems evident in the natural settling areas within Jacobs River Estuary, and following preliminary synoptic fine scale assessment undertaken in the Waihopai Arm of New River Estuary in February 2011 which indicated significant degradation in these settling areas (see Robertson and Stevens 2011a), detailed fine scale assessment of the natural settling areas within both Aparima and Pourakino Arms was undertaken using the NEMP approach (Robertson et al. 2002). Sampling was undertaken in late January 2012 and results are presented in the current report.

Jacobs River Estuary is a moderate sized "tidal lagoon" type estuary (area 720ha) situated at the confluence of the Pourakino and Aparima Rivers that discharges to the sea at Riverton. The estuary is shallow (mean depth ~2m) and has a mixture of poorly flushed and well flushed areas. It drains a primarily agricultural catchment bordered by a mix of vegetation and landuses (predominantly grazed pasture and urban). Human use of the estuary is high and is used for walking, shellfish collecting, boating, fishing, duck shooting, bird watching, bathing, and white-baiting.

The estuary has extensive sand and mudflats, seagrass, and saltmarsh areas. Habitat diversity is moderate with tidal flats and saltmarsh providing important habitat for native fish, birdlife and tidal flat organisms. However, as a consequence of historical drainage, extensive weed growth, intensive farming in the catchment, and the grazing of margins, poorly flushed parts of the estuary are now relatively vulnerable to eutrophication and sedimentation. This is exacerbated in the Pourakino Arm by the natural constriction of the Narrows" (see Figure 1.) Catchment development is evident in moderately degraded water quality entering the estuary, common nuisance blooms of macroalgae (*Ulva* and *Gracilaria*), and accumulations of deep soft muds. As a consequence, the estuary has several very eutrophic arms.

Therefore, it has been recommended that management actions be taken to improve the situation.

1. Introduction (Continued)

Table 1. Summary of the major issues affecting most NZ estuaries.

Table 2. Summary of the broad and fine scale EMP indicators (shading signifies indicators used in the fine scale monitoring assessments)**.**

2. METHODS

Fine scale monitoring

Fine scale monitoring is based on the methods described in the NEMP (Robertson et al. 2002) and provides detailed information on the condition of the estuary. Using the outputs of the broad scale habitat mapping, representative sampling sites (usually one or two per estuary, or three or four for larger estuaries) are selected and samples collected and analysed for physical, chemical and biological variables.

In addition to the existing three fine scale sites located in the dominant sandy intertidal flats of Jacobs River Estuary, two additional fine scale sampling sites (Figure 1, Appendix 2) were selected in mid-low water habitat within the eutrophic natural settling areas of the estuary (areas with abundant macroalgal growth and muddy anoxic, sulphide-rich sediment). These eutrophic areas represent ~1/3rd of the total estuary habitat. At each eutrophic site, a 20m x 8m area in the lower intertidal was marked out and divided into 10 equal sized plots. Within each plot, a random position was defined, and the following sampling undertaken:

Physical and chemical analyses

- One core was collected to a depth of at least 100mm and photographed alongside a ruler and a corresponding label. Colour and texture were described and average redox potential discontinuity (RPD) depth (i.e. depth to light grey/black anoxic layer) recorded.
- At each site, three samples (2 a composite from 4 plots, and 1 from 2 plots) of the top 20mm of sediment (each approx. 250gms) were collected adjacent to each core. All samples were kept in a chillybin in the field.
- Chilled samples were sent to R.J. Hill Laboratories for analysis of the following (details in Appendix 1):
	- Grain size/Particle size distribution (% mud, sand, gravel).
	- Nutrients total nitrogen (TN), total phosphorus (TP), and total organic carbon (TOC).
	- Trace metal contaminants (total recoverable Cd, Cr, Cu, Ni, Pb, Zn). Analyses were based on whole sample fractions which are not normalised to allow direct comparison with the Australian and New Zealand Guidelines for Fresh and Marine Water Quality (ANZECC 2000).
- Samples were tracked using standard Chain of Custody forms and results checked and transferred electronically to avoid transcription errors.
- Photographs were taken to record the general site appearance.

Epifauna (surface-dwelling animals)

Epifauna were assessed from one random $0.25m^2$ quadrat within each of the ten plots. All animals observed on the sediment surface were identified and counted, and any visible microalgal mat development noted. The species, abundance and related descriptive information were recorded on waterproof field sheets containing a checklist of expected species. Photographs of quadrats were taken and archived for future reference.

Infauna (animals within sediments)

- One randomly placed sediment core was taken from each of 10 plots using a 130mm diameter (area = 0.0133 m²) PVC tube.
- The core tube was manually driven 150mm into the sediments, removed with the core intact and inverted into a labelled plastic bag.
- Once all replicates had been collected at a site, the plastic bags were transported to a nearby source of seawater and the contents of the core were washed through a 0.5mm nylon mesh bag. The infauna remaining were carefully emptied into a plastic container with a waterproof label and preserved in 70% isopropyl alcohol - seawater solution.
- The samples were then transported to a commercial laboratory for counting and identification (Gary Stephenson, Coastal Marine Ecology Consultants, Appendix 1).

2. Methods (Continued)

1. Pourakino Arm eutrophic Site D and fine scale Site C. 2. Aparima Arm eutrophic Site E.

Figure 1. Location of fine scale monitoring sites in Jacobs River Estuary (Photo ES).

2. Methods (Continued)

Sedimentation Rate

Determining the sedimentation rate from now and into the future involves a simple method of measuring how much sediment builds up over buried plates over time. Once a plate has been buried, levelled, and the elevation measured, probes are pushed into the sediment until they hit the plate and the penetration depth is measured. A number of measurements on each plate are averaged to account for irregular sediment surfaces, and a number of plates are buried to account for small scale variance. Locations (Figure 1) and methods for deployment are presented in Robertson and Stevens 2011). In the future, these depths will be measured every 1-5 years and, over the long term, will provide a measure of rate of sedimentation in representative parts of the estuary.

Central Basin sedimentation rate site in Jan. 2012 Pourakino Arm sedimentation rate site in Jan. 2012.

Condition Ratings

A series of interim fine scale estuary "condition ratings" (presented below) have been proposed for Jacobs River Estuary (based on the ratings developed for Southland's estuaries - e.g. Robertson & Stevens 2006). The ratings are based on a review of estuary monitoring data, guideline criteria, and expert opinion. They are designed to be used in combination with each other (usually involving expert input) when evaluating overall estuary condition and deciding on appropriate management. The condition ratings include an "early warning trigger" to highlight rapid or unexpected change, and each rating has a recommended monitoring and management response. In most cases initial management is to further assess an issue and consider what response actions may be appropriate (e.g. develop an Evaluation and Response Plan - ERP).

2. Methods (Continued)

3. RESULTS AND DISCUSSION

A summary of the 29 January 2012 eutrophic zone monitoring results is presented alongside the long term fine scale monitoring results (2003-2006 baseline and 2011 results - Robertson and Stevens 2011) in Table 3. Sedimentation rate monitoring results are presented in Table 4, and detailed macroinvertebrate results in Table 5. The results and discussion section is divided into three subsections based on the key estuary problems that the fine scale monitoring is addressing: sedimentation, eutrophication, and toxicity.

Sedimentation

Accelerated soil erosion from developed catchments is a major issue for tidal lagoon estuaries in NZ as they form a sink for fine suspended sediments. NZ estuaries are particularly sensitive to increased muddiness given the facts that they are generally sand dominated, have a diverse and healthy biology, and have a short history of catchment development. Increased muddiness results in reduced sediment oxygenation, production of toxic sulphides, increased nuisance macroalgal growth, and a shift towards a degraded invertebrate and plant community. Such a change reduces feeding grounds and habitat for bird and fish species. Unless the input of fine sediment is reduced to a level within the assimilative capacity of the estuary, then the estuary will rapidly infill, high value habitat will be lost, and their value for fish, birdlife and humans be greatly reduced.

Sediments containing high mud content (i.e. around 30% mud with a grain size <63μm) are now typical in NZ estuaries that drain developed catchments. In such mud-impacted estuaries, the muds are generally concentrated in areas that experience low energy tidal currents and waves i.e. the intertidal margins of the upper reaches of estuaries (e.g. Waihopai Arm, New River Estuary sites W,E, F, Jacobs River Estuary sites D, E - Figure 2).

Table 3. Physical, chemical and macrofauna results (as means) for main basin Jacobs River Estuary sites (2003-2011) and eutrophic arms (2012).

Site		RPD	T ₀ C	Mud Sand		Gravel	C _d	Cr	Pb Ni Cu		Zn	TN	TP	Abundance		Mean Species No.	
		cm		%						mg/kg					$No./m^2$	per core	per site
Sites located in the relatively well flushed central basin of Jacobs River Estuary																	
2003	JRE Fine Scale A	be provided. unable to \mathbf{I} Cawthron Data collected by	0.9	1.0	98.8	0.22	$0.48*$	14.70	24.0	8.36	12.08	73.1	121	495	4508	8	21
	JRE Fine Scale B		0.4	2.8	96.5	0.71	$0.19*$	11.90	6.19	4.44	5.90	17.3	147	203	3870	12	24
	JRE Fine Scale C		0.7	4.0	95.3	0.80	$0.11*$	10.43	5.57	3.87	4.72	17.4	222	276	5453	10	20
2004	JRE Fine Scale A		1.1	0.8	98.8	0.45	< 1.0	10.27	23.2	2.90	11.10	83.1	149	464	7335	14	23
	JRE Fine Scale B		0.6	3.9	95.1	1.05	< 1.0	6.93	5.8	1.64	5.57	50.7	252	198	6780	15	28
	JRE Fine Scale C		0.8	6.4	92.8	0.81	< 1.1	7.16	4.91	1.61	4.87	51.6	249	267	9330	12	22
2005	JRE Fine Scale A		1.0	4.9	94.7	0.46	< 0.05	9.87	23.7	11.90	3.22	52.3	255	508	6105	12	21
	JRE Fine Scale B		0.6	6.2	93.0	0.75	< 0.05	8.21	6.61	6.40	1.85	19.2	262	248	5693	15	32
	JRE Fine Scale C		0.7	8.7	90.7	0.59	< 0.05	7.89	5.66	5.40	1.81	20.0	358	336	7013	13	24
2006	JRE Fine Scale A		1.0	1.5	98.0	0.50	0.10	8.87	21.67	11.33	2.90	56.3	140	577	6188	13	26
	JRE Fine Scale B		0.9	7.7	92.1	0.20	0.10	7.13	6.47	6.00	1.73	19.0	250	219	5828	16	26
	JRE Fine Scale C		1.0	13.6	85.9	0.57	0.10	7.47	6.47	5.33	2.03	21.3	333	389	4163	10	22
2011	JRE Fine Scale A	$\mathbf{0}$	0.1	6.9	92.9	0.23	0.07	10.60	25.0	13.20	3.43	58.3	< 500	500	9421	11	24
	JRE Fine Scale B	$\mathbf{1}$	0.2	4.8	94.8	0.37	0.02	7.03	6.17	5.73	1.71	16.8	< 500	204	6098	17	30
	JRE Fine Scale C	1	0.2	4.7	94.6	0.63	0.02	6.97	5.13	4.90	1.65	16.8	< 500	233	18715	15	23
						Sites located in relatively poorly flushed sheltered arms of Jacobs River Estuary											
2012	JRED (Pourakino)	$\mathbf{0}$	1.2	52.4	46.9	1.1	0.12	17.2	30.7	15.4	5.4	65.0	1433	640	18120	10	17
	JRE E (Aparima)	$0 - 0.5$	2.7	63.0	36.0	1.0	0.09	22.3	25.7	16.2	6.2	54.3	3133	823	11272	8	16
	*denotes unreliable chemical results																

Mud dominated sediments in the Pourakino Arm.

Figure 3. Grain size, Jacobs River Estuary.

Figure 4. Cumulative change in sediment levels over buried plates in Jacobs River Estuary, 2011 to 2012.

In contrast, the main intertidal flats of developed estuaries (e.g. New River Estuary sites B, C, D, and Jacobs River Estuary sites A, B, C - Figure 2) are usually characterised by sandy sediments low in mud content (2-10% mud) reflecting their exposure to wind-wave disturbance. In order to assess sedimentation in the Pourakino and Aparima Arms of Jacobs River Estuary, a number of indicators have been used: grain size, sedimentation rate, and presence of mud tolerant invertebrates.

Grain Size

Grain size (% mud, sand, gravel) measurements provide a good indication of the muddiness of a particular site. The monitoring results for all Jacobs River Estuary sites (Figure 3) show that the Pourakino and Aparima eutrophic zone sediments (Sites D and E) were dominated by mud (52-63% mud), whereas the main intertidal basin sites were dominated by sands (>90% sand in all years).

Compared with fine scale sites in other tidal lagoon type estuaries in Southland, the Pourakino and Aparima Arm mud contents were very high (Figure 2). Such findings are not unexpected given the intensively developed pastoral nature of the catchment and the elevated sediment yields expected from such landuse areas. Estimated areal sediment loads to Jacobs River Estuary are 18.3g.m⁻².d⁻¹ (based on CLUES model estimates), which is nearly twice the areal load to the nearby similarly affected New River Estuary. Both estuaries have extensive muddy, anoxic, eutrophic zones within their borders.

Sedimentation Rate

Table 4 presents the January 2012 sedimentation rate monitoring results for the 12 plates buried in Jacobs River Estuary, with summary data from 2011-2012 presented in Figure 4.

Although the results are very much preliminary, reflecting only 1 year of monitoring, sediment rates in the main central basin of the estuary (Sites A and B) were rated in the "very low" category. From Feb. 2011 to Jan. 2012, both sites showed a decrease in sediment which is attributed to wave driven movement of surface sediments common in the main estuary basin. At Site C, in the muddy, eutrophic and macroalgae covered upper Pourakino Arm, there was a "moderate" increase in sediment recorded.

The results reflect the obvious differences between the sites. The well flushed main basin sites remain dominated by sands, with muds not accumulating on the estuary surface. In contrast, large parts of the sheltered arms of the estuary are smothered in deep soft, sulphide-rich muds and excessive macroalgal growth. The source of fine muds is almost certainly from the surrounding Pourakino and Aparima catchments rather than the sea. While ongoing monitoring of sedimentation rates will measure changes into the future, the prevailing conditions indicate the assimilative capacity of the settling basins within the estuary are currently being exceeded and that there has already been a significant input of muddy sediments into the estuary.

Table 4. Sedimentation rate monitoring results, Jacobs River Estuary, February 2011 - January 2012.

Thick beds of *Gracilaria* **that contribute to trapping and deposition of fine muds in the Upper Pourakino Arm.**

Macro-invertebrate Community

Sediment mud content is a major determinant of the structure of the benthic invertebrate community. This section examines this relationship in Jacobs River Estuary in three steps by:

- 1. Comparing the mean abundance of major macroinvertebrate groups, and the number of species in the surface and subsurface feeding types, with other estuary sites to see if there are any major differences (Figures 5 and 6).
- 2. Using the response of typical NZ estuarine macro-invertebrates to increasing mud content (Gibbs and Hewitt 2004) to assess the mud tolerance of the Jacobs River Estuary macro-invertebrate community over the 11 years of monitoring (Figure 7).
- 3. Using multivariate techniques to explore whether the macro-invertebrate communities at each of the Pourakino and Aparima eutrophic sites differ from the other estuary sites (Figure 8).

As previously explained, ten core samples were analysed at each of the gross eutrophic sites, Pourakino Arm (Sites D) and Aparima Arm (Site E), for the presence of macroinvertebrates. The samples were taken from the dominant habitat type in each arm, i.e. muddy, anoxic, sulphide rich sediments overlain with a thick layer of partially decaying macroalgae.

The 2012 results (Appendix 3) indicate a community that was dominated by crustacea, gastropods and bivalves, which was different to the community composition at the other fine scale sites measured previously in Jacobs River Estuary (Figure 5), except for Site C in 2011. These differences are attributed to the poor sediment conditions at the eutrophic sites and the fact that they were overlain with a thick layer of decaying algae. The proximity to large areas of macroalgal cover in the Pourakino Arm was the likely explanation for the large numbers of gastropods (primarily *Potamopyrgus* sp.) found at Site C in 2011.

Figure 5. Mean abundance of major infauna groups, Jacobs River Estuary, 2003-2012.

Such conditions favour surface feeding organisms (particularly crustacea, gastropods, bivalves and some polychaetes), rather than subsurface animals that feed on deposits within the sediments (Figure 6). In particular, the gross eutrophic sites included the following taxa in relatively large numbers:

- The tube-dwelling crustacean amphipod *Paracorophium excavatum*, which is the dominant corophioid amphipod in the South Island. *Paracorophium* is well-known as a major primary coloniser (and hence indicator) of disturbed estuarine intertidal flats (Ford et al. 1999). Examples of common disturbances are: macroalgal mats settling on the tidal flats as a result of coastal eutrophication, and mud deposition after mobilisation of fine sediments from exposed soil surfaces in the catchment. In these situations, *Paracorophium* can become very abundant and, through its burrowing activities, increases oxygen exchange, which in turn helps mitigate the effect of the disturbance.
- Large numbers of other unidentified amphipods associated with the surface layer.
- The small deposit-feeding bivalve Arthritica bifurca, that prefers living in muddy-sand habitats, but is tolerant to a broad range of mud contents. These bivalves do not grow more than 5mm in size and feed at a depth of greater than 2cm.
- • The small native estuarine snails *Potamopyrgus sp.* that require brackish conditions for survival. They feed on decomposing animal and plant matter, bacteria and algae, and while tolerant of muds, are intolerant of anoxic sediments.
- • The surface deposit feeding spionid polychaete *Scolecolepides benhami*. This spionid is very tolerant of mud, fluctuating salinities, organic enrichment and toxicants (e.g. heavy metals). It is rarely absent in sandy/mud estuaries, often occurring in a dense zone high on the shore, although large adults tend to occur further down towards the low water mark.

Figure 6. Mean number of species within surface and subsurface feeding groups, Jacobs River Estuary, 2003-2012.

These taxa are widely acknowledged to respond in a characteristic manner to disturbance and are sensitive to pollution. Studies on the response of invertebrates to increased clay/silt sediments (Norkko et al., 2001, and the results from a wide range of NZ estuaries (Robertson and Stevens, in prep), have identified all the dominant taxa as tolerant to, or preferring, increased mud content (Table 5: MUD Groups 3-5).

When compared to the other less muddy fine scale monitoring sites in the main basin of Jacobs River Estuary (Sites A and B), the macroinvertebrate mud tolerance rating was in the fair range at the gross eutrophic sites, indicating a community dominated by mud tolerant taxa (Figure 7). The similarity of Site C to the eutrophic sites likely reflects the higher mud content evident at Site C compared to Sites A and B.

Multivariate techniques were also used to explore whether the macroinvertebrate communities at the gross eutrophic sites differed from the less disturbed fine scale sites in Jacobs River Estuary (Figure 8). Figure 8 shows that the results of the multivariate analysis (NMDS Plot) clearly portray the difference in the benthic invertebrate communities between the cleaner, less disturbed, low mud content sites (Sites A,B, and C) and the muddy, eutrophic sites (Sites D and E). Data from other studies (Robertson and Stevens 2012a) showed that macrofauna communities were more diverse and abundant where there was a thick macroalgal layer over anoxic sulphide rich sediments (as at the Jacobs River sites D and E), compared to the sites with thick anoxic muds and no macroalgal layer.

Figure 8. NMDS plot showing the relationship among mean samples in terms of similarity in macro-invertebrate community composition for Jacobs River Estuary Sites A, B, C, D, and E and for 2003-2012. The plot shows the mean of each of the 10 (or 12 in 2003) replicate samples for each site and is based on Bray Curtis dissimilarity and fourth root transformed data.

The approach involves multivariate data analysis methods, in this case non-metric multidimensional scaling (NMDS) using PRIMER version 6.1.10. The analysis basically plots the site, year and abundance data for each species as points on a distance-based matrix (a scatterplot ordination diagram). Points clustered together are considered similar, with the distance between points and clusters reflecting the extent of the differences. The interpretation of the ordination diagram depends on how good a representation it is of actual dissimilarities i.e. how low the calculated stress value is. Stress values greater than 0.3 indicate that the configuration is no better than arbitrary, and we should not try and interpret configurations unless stress values are less than 0.2.

muddy macroalgal cover.

Aparima Arm Site E showing thin macroalgal cover overlaying surface muds.

Pourakino Arm Site D showing thick **Number 20 Aparima Arm Site E showing thin mac-** Luxuriant subtidal growth in the Aparima River.

Figure 10. Total organic carbon (median, upper and lower quartiles, range) at intertidal sites, 2003-2012.

tiles, range) at intertidal sites, 2003-2012.

Eutrophication

Excessive organic input, sourced either from outside the estuary or growing within it in response to high nutrient loads, is a principal cause of physical and chemical degradation and of faunal change in estuarine and nearshore benthic environments. In tidal lagoon estuaries like Jacobs River, as organic input to the sediment increases the sediments become deoxygenated, nuisance algal growth becomes abundant, the number of suspension-feeders (e.g. bivalves and certain polychaetes) declines, and depositfeeders (e.g. opportunistic polychaetes) increase (Pearson and Rosenberg 1978). The primary fine scale indicators of eutrophication are grain size, RPD depth, sediment organic matter, nitrogen and phosphorus concentrations, and the community structure of certain sediment-dwelling animals. The broad scale indicators are the percentages of the estuary covered by macroalgae and soft muds (Stevens and Robertson, 2008).

Redox Potential Discontinuity (RPD)

The depth of the RPD boundary provides an indication of the level of sediment oxygenation. The results (Figure 9) showed the 2012 RPD depth at the Pourakino and Aparima Arm eutrophic sites was at, or very near, the surface (0- 0.5cm), reflecting very poorly oxygenated sediments. These RPD ratings were similar to those measured at the main basin fine scale sites in 2011. Such shallow RPD values fit the "poor" condition rating (see Section 2), with the benthic invertebrate community likely to be dominated by a few pollution-tolerant species that live near the surface.

Total Organic Carbon and Nutrients

The concentrations of sediment nutrients (total nitrogen - TN and phosphorus - TP) and organic matter (total organic carbon - TOC) also provide valuable trophic state information. In particular, if concentrations are elevated, and eutrophication symptoms are present (i.e. shallow RPD, excessive algal growth, low biotic index), then N, P and TOC concentrations provide a good indication that loadings are exceeding the assimilative capacity of the estuary. However, a low TOC, TN, or TP concentration does not in itself indicate an absence of eutrophication symptoms as the estuary, or part of an estuary, may have reached a eutrophic condition and exhausted the nutrient supply. Obviously, the latter case is likely to better respond to input load reduction than the former.

In relation to the eutrophic arms of the Jacobs River Estuary (Sites D and E), the results (Figures 10-12) indicate elevated concentrations of TOC, TP and TN compared with fine scale sites in less eutrophic parts of the estuary. Note, a changed TN method in 2012 is likely to underestimate TN compared to previous values by 10-40%. Combined with the very high macroalgal cover (Site D=100%, Site E (10- 100%) recorded in 2012, and the shallow RPD depths, these results confirm the eutrophic nature of these estuary arms and the oversupply of sediment nutrients in the area.

Macro-invertebrate Organic Enrichment Index

The benthic invertebrate organic enrichment tolerance ratings for the eutrophic arm sites in Jacobs River Estuary (JR D and JR E) were in the "low" category (Figure 13). However, because the sediment cores had only a very few sediment dwelling species (Figure 6), Type I "very sensitive" organisms were only present at the surface, and many of the surface species (particularly Amphipoda) that dominated the eutrophic samples have not yet been ascribed tolerances to organic enrichment (Table 5), the AMBI is considered to currently under-represent the extent of sediment degradation at the Jacobs River eutrophic sites.

Instead, the "low" rating (indicating only slight organic enrichment) is attributed to the presence of a thick surface macroalgal layer with a relatively low mud content. This supports a community of surface feeding organisms with varying tolerances to organic enrichment as they are not constantly exposed to the degraded (anoxic and sulphide-rich) conditions in the underlying sediment. Beneath this surface macroalgal layer, few animals are able to survive.

The impact caused to the sediment community by enrichment is most apparent when comparing species composition over time from JR C in the Pourakino Arm (Figure 1). Here, the shift to increased eutrophication, particularly from 2006 onwards, has seen a classical corresponding pollution response (e.g. Warwick 1986) of increased abundance of smaller opportunistic species and decreased community biomass and diversity (Figure 5). This site, along with the eutrophic site JR D, now show clear signs of ecological stress.

The presence of significant adverse sediment impacts in Jacobs River Estuary is supported by data from New River Estuary (Robertson and Stevens 2012a) which showed that where very anoxic muds are present without an oxygenated surface macroalgal layer, the community had a "very high" AMBI rating, indicating a significantly degraded macroinvertebrate community. Although areas with such conditions were present in Jacobs River Estuary, the 2012 monitoring focused on the dominant habitat of the eutrophic arms which were areas with surface macroalgae.

Figure 13. Organic enrichment macro-invertebrate rating, Jacobs River Estuary, 2003-2012.

Pourakino Arm Aparima Arm, thick macroalgal cover (centre) and site patchiness (right)

Toxicity

Heavy metals (Cd, Cr, Cu, Ni, Pb, Zn), used as an indicator of potential toxicants, were at moderate concentrations at the Pourakino and Aparima Arm eutrophic sites D and E, with all non-normalised values below the ANZECC (2000) ISQG-Low trigger values (Figure 14). However, these concentrations were generally higher than those measured at the main basin sites during 2003-2011. Such conditions indicate a moderate accumulation of heavy metals in the sediments of the Pourakino and Aparima Arms of the estuary.

Pourakino Arm Site D showing deep soft anoxic muds present beneath macroalgae.

4. Conclusio ns

The 2012 results indicate that the poorly flushed Pourakino and Aparima Arms of the estuary are excessively muddy, have high nutrients and nuisance macroalgal growths, and very poor sediment oxygenation. As a result, the macroinvertebrate community is dominated by surface feeding species that are tolerant of such poor conditions. Such conditions limit the food availability for fish and birdlife, and mean the capacity of the estuary to assimilate nutrient and sediment loads from the catchment is exceeded in the upper arm settlement zones. Toxicity (indicated by heavy metals) was low, but higher than measured at other fine scale sites.

Issues identified in other monitoring studies of Jacobs River Estuary include; loss of high value habitat, excessive muddiness in the sheltered estuary arms, disease risk associated with shellfish consumption and bathing, and toxicity near urban stormwater drains.

The results confirm the presence of significant areas of gross eutrophic conditions in sheltered estuary arms. Results are as summarised below, and compared with earlier results for sites in the main body of the estuary, as follows:

It is expected that these extreme conditions that occur at the Pourakino and Aparima sites are due to the sheltered nature of these arms and their propensity to act as natural settling areas for fine sediment and macroalgae sourced from both within and outside the estuary. Because these areas provide an early warning of siltation and eutrophication problems to the wider estuary, and are already showing significant and rapid degradation, it is recommended that the Pourakino and Aparima Arm sites be included in the long term monitoring estuary programme.

5. Monitorin g

Jacobs River Estuary has been identified by Environment Southland as a high priority for monitoring, and is a key part of their coastal monitoring programme being undertaken in a staged manner throughout the Southland region. The future monitoring recommendations are outlined as follows:

Fine Scale Monitoring.

Monitor Pourakino and Aparima Arm Sites D and E in February 2013, 2014 and again in February 2016 when the 5 yearly fine scale trend monitoring falls due.

Macroalgal and Seagrass Monitoring.

Continue with the programme of annual broad scale mapping of macroalgae. Next monitoring due in February 2013. In addition, in order to assess changes in seagrass cover, it is recommended that seagrass cover be monitored annually in priority areas in tandem with the macroalgal monitoring.

Broad Scale Habitat Mapping.

Continue with the programme of 5 yearly broad scale habitat mapping. Next monitoring due in February/March 2013.

Sedimentation Rate Monitoring.

Because sedimentation is a priority issue in the estuary it is recommended that sediment plate depths be measured annually.

6. MANAGEMENT

Eutrophication and sedimentation have been identified as major issues in Jacobs River Estuary since at least 2007-2008 (Robertson and Stevens 2008, Stevens and Robertson 2008), as has been the case for several other Southland estuaries (e.g. Jacobs River, Waimatuku and Waituna Lagoon).

To address these issues, it is recommended that catchment nutrient and sediment guideline criteria be developed for each estuary type in Southland in a prioritised fashion (e.g. Robertson and Stevens 2011). New River Estuary is recommended as the first priority, with the results being applied to Jacobs River Estuary to quickly assess the extent to which current catchment loads are likely to meet likely guideline criteria. If the catchment inputs are at or below guideline criteria, the estuary is expected to flourish and provide sustainable human use and ecological values in the long term. If catchment loads exceed the estuary's assimilative capacity, it will continue to degrade.

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8. Referen ces

ANZECC. 2000. Australian and New Zealand guidelines for fresh and marine water quality. Australian and New Zealand Environment and Conservation Council, Agriculture and Resource Management Council of Australia and New Zealand.

- *Borja, A., Franco, J., and Perez, V. 2000. A marine biotic index to establish the ecological quality of soft-bottom benthos within European estuarine and coastal environments. Mar. Poll. Bull. 40, 1100–1114.*
- *Borja, A. and Muxika, H. 2005. Guidelines for the use of AMBI (AZTI's Marine Biotic Index) in the assessment of the benthic ecological quality. Marine Pollution Bulletin 50: 787-789.*
- *Environment Southland. 2012. Water Quality Methodology for Southland Water 2010 Report. Environment Southland, Invercargill. Publication number 2012-05.*
- *Ford, R.B., Thrush, S.F. and Probert, P.K. 1999. Macrobenthic colonisation of disturbances on an intertidal sandflat: the influence of season and buried algae Marine Ecology Progress Series 191: 163-174.*
- *Gibbs, M. and Hewitt, J. 2004. Effects of sedimentation on macrofaunal communities: a synthesis of research studies for ARC. Technical Paper 264. NIWA Client Report: HAM2004-060.*
- *Jørgensen, N. and Revsbech, N.P. 1985. Diffusive boundary layers and the oxygen uptake of sediments and detritus. Limnology and Oceanography 30:111-122.*
- *Norkko, A., Talman, S., Ellis, J., Nicholls, P. and Thrush, S. 2001. Macrofaunal sensitivity to fine sediments in the Whitford embayment. NIWA Client Report ARC01266/2 prepared for Auckland Regional Council.*
- *Pearson, T.H. and. Rosenberg, R. 1978. Macrobenthic succession in relation to organic enrichment and pollution of the marine environment. Oceangraph and Marine Biology Annual Review 16, 229–311.*
- *Robertson, B.M., Gillespie, P.A., Asher, R.A., Frisk, S., Keeley, N.B., Hopkins, G.A., Thompson, S.J., Tuckey, B.J. 2002. Estuarine Environmental Assessment and Monitoring: A National Protocol. Part A. Development, Part B. Appendices, and Part C. Application. Prepared for supporting Councils and the Ministry for the Environment, Sustainable Management Fund Contract No. 5096. Part A. 93p. Part B. 159p. Part C. 40p plus field sheets.*
- *Robertson B.M., Tuckey B.J., and Robertson B. 2003. Broad scale mapping of Jacobs River Estuary intertidal habitats. Prepared for Environment Southland.*
- *Robertson, B.M. and Stevens, L.M. 2006. Southland Estuaries State of Environment Report 2001-2006. Prepared for Environment Southland. 45p plus appendices.*
- *Robertson, B.M. and Stevens, L.M. 2008. Southland Coast Te Waewae Bay to the Catlins, habitat mapping, risk assessment and monitoring recommendations. Report prepared for Environment Southland. 165p.*
- *Robertson, B.M. and Stevens, L.M. 2011. Jacobs River Estuary Fine Scale Monitoring 2010/11. Report prepared by Wriggle Coastal Management for Environment Southland. 34p.*
- *Robertson, B.M. and Stevens, L.M. 2011a. Waihopai Arm New River Estuary Preliminary Synoptic Assessment 2010/11. Report prepared by Wriggle Coastal Management for Environment Southland. 16p.*
- *Robertson, B.M. and Stevens, L.M. 2012a. New River Estuary, Fine scale monitoring of highly eutrophic arms 2011/2012. Report prepared by Wriggle Coastal Management for Environment Southland. 27p.*
- *Robertson, B.M. and Stevens, L.M. 2012b. Guidance Document: Nutrient Load Criteria to Limit Eutrophication in Three Typical New Zealand Estuary Types - ICOLL's, Tidal Lagoon, and Tidal River Estuaries. Report prepared by Wriggle Coastal Management for Environment Southland. 7p.*
- *Stevens, L.M. and Robertson, B.M. 2008. Jacobs River Estuary. Broad Scale Habitat Mapping 2007/08. Report prepared by Wriggle Coastal Management for Environment Southland. 31p.*
- *Stevens, L.M. and Robertson, B.M. 2010. Jacobs River Estuary. Macroalgal Monitoring 2009/10. Report prepared by Wriggle Coastal Management for Environment Southland. 7p.*
- *Stevens, L.M. and Robertson, B.M. 2011. Jacobs River Estuary. Macroalgal Monitoring 2010/11. Report prepared by Wriggle Coastal Management for Environment Southland. 6p.*
- *Stevens, L.M. and Robertson, B.M. 2012. Jacobs River Estuary. Macroalgal Monitoring 2011/12. Report prepared by Wriggle Coastal Management for Environment Southland. 6p.*
- *Thrush, S.F., Hewitt, J.E., Norkko, A., Nicholls, P.E., Funnell, G.A., and Ellis, J.I. 2003. Habitat change in estuaries: predicting broad-scale responses of intertidal macrofauna to sediment mud content. Marine Ecology Progress Series 263:101-112.*
- *Thrush, S.F., Hewitt, J.E., Gibb, M., Lundquist, C., and Norkko, A. 2006. Functional role of large organisms in intertidal communities: Community effects and ecosystem function. Ecosystems 9: 1029-1040.*
- *Warwick, R.M. 1986. A new method for detecting pollution effects on marine macrobenthic communities. Marine Biology. 92: 557–562.*

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Appendix 1. Analytical Methods

***** Coastal Marine Ecology Consultants (established in 1990) specialises in coastal soft-shore and inner continental shelf soft-bottom benthic ecology. Principal, Gary Stephenson (BSc Zoology) has worked as a marine biologist for more than 25 years, including 13 years with the former New Zealand Oceanographic Institute, DSIR. Coastal Marine Ecology Consultants holds an extensive reference collection of macroinvertebrates from estuaries and soft-shores throughout New Zealand. New material is compared with these to maintain consistency in identifications, and where necessary specimens are referred to taxonomists in organisations such as NIWA and Te Papa Tongarewa Museum of New Zealand for identification or cross-checking.

Appendix 2. 2012 Detailed Results

Station Locations

Physical and Chemical Results for Jacobs River Estuary (Sites D and E), 29 January 2012.

***** composite samples

APPENDIX 2. 2012 DETAILED RESULTS (CONTINUED)

Epifauna and macroalgal cover within 0.25m2 quadrats, Jacobs River Estuary Sites D and E, 29 Jan. 2012.

Macroinvertebrate Infauna for Jacobs River Estuary (Sites D and E), 29 January 2012 - numbers per core.

Appendix 3. Infauna Characteristics

Wriggle coastalmanagement28

NA=Not Allocated

1 = SS, strong sand preference.

 $2 = S$, sand preference.

 $3 =$ I, prefers some mud but not high percentages.

- $4 = M$, mud preference.
- 5 = MM, strong mud preference.

***** AMBI Sensitivity to Organic Enrichment Groupings (from Borja et al. 2000)

Group I. Species very sensitive to organic enrichment and present under unpolluted conditions (initial state). They include the specialist carnivores and some deposit-feeding tubicolous polychaetes.

Group II. Species indifferent to enrichment, always present in low densities with non-significant variations with time (from initial state, to slight unbalance). These include suspension feeders, less selective carnivores and scavengers.

Group III. Species tolerant to excess organic matter enrichment. These species may occur under normal conditions, but their populations are stimulated by organic enrichment (slight unbalance situations). They are surface deposit-feeding species, as tubicolous spionids.

Group IV. Second-order opportunistic species (slight to pronounced unbalanced situations). Mainly small sized polychaetes: subsurface deposit-feeders, such as cirratulids.

Group V. First-order opportunistic species (pronounced unbalanced situations). These are deposit-feeders, which proliferate in reduced sediments.

The distribution of these ecological groups, according to their sensitivity to pollution stress, provides a Biotic Index with 5 levels, from 0 to 6.

